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1 Long-term persistence of pioneer species in tropical forest soil seed banks

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18

19

19 **Abstract**

20 In tropical forests, pioneer species regenerate from seeds dispersed directly into canopy
21 gaps, and from seeds that persisted in soil seed banks before gap formation. However,
22 life-history models suggest that selection for long-term persistence of seeds in soil should
23 be weak, as persistence incurs a fitness cost resulting from prolonged generation time.
24 We use a carbon dating technique to provide the first direct measurements of seed
25 persistence in undisturbed tropical forest seed banks. We show that seeds germinate
26 successfully from surface soil microsites up to 38 years after dispersal. Decades-long
27 persistence may be common in pioneers with relatively large mass, and appears to be
28 unrelated to specific regeneration requirements. In *Croton billbergianus*, a sub-canopy
29 tree that recruits in abundant small gaps, long-term persistence is associated with short-
30 distance ballistic seed dispersal. In *Trema micrantha*, a canopy tree with widespread
31 dispersal, persistence is associated with a requirement for large gaps that form
32 infrequently in old-growth forest.

33

33 Introduction

34

35 Up to 15 % of tree species in old-growth tropical forests have been classified as
36 pioneers that require high light conditions for successful seedling recruitment (Hubbell et
37 al. 1999, Molino and Sabatier 2001). In most forests these conditions are found in treefall
38 gaps that occur infrequently, at largely unpredictable locations, and that usually only
39 persist for only a few years before canopy closure (Hartshorn 1990, Young and Hubbell
40 1991, Schnitzer et al. 2000). As a consequence, pioneer life-histories are usually
41 characterized by traits associated with a high colonization ability *i.e.* prolific seed
42 production, high seed dispersability, and disturbance-cued germination from soil seed
43 banks (Whitmore 1983, Swaine and Whitmore 1988).

44

45 Despite presumably strong selection for traits that favor the colonization and
46 occupation of gaps, pioneer life-histories are remarkably diverse. On Barro Colorado
47 Island (BCI), Panama, pioneer species vary by over four orders of magnitude in seed
48 mass (Dalling et al. 1998b), and at least two-fold in seedling relative growth rate (Dalling
49 et al. 2004). This trait diversity is maintained by trade-offs that equalize recruitment
50 success (Dalling and Burslem 2005). Seed size variation is maintained by a trade-off
51 between colonization success, selecting for large numbers of small seeds, and emergence
52 and establishment success, selecting for large seed mass (Dalling and Hubbell 2002,
53 Coomes and Grubb 2003). Similarly, variation in seedling relative growth rate can be
54 explained by an inverse relationship between growth and survival rate (Brokaw 1987,
55 Hubbell and Foster 1992, Kitajima 1994, Dalling et al. 1998b).

Trade-offs affecting establishment and growth, however, cannot account for additional interspecific variation observed in seed dispersal ability. Among the pioneers of BCI, median dispersal distances estimated from seed trap data vary from <1 m for some species with dehiscent fruit or ballistic dispersal, to > 60 m for some wind-dispersed tree species (Dalling et al. 2002). Seed burial experiments suggest that significant variation also exists in seed persistence time. While detailed demographic studies of common, small-seeded pioneers have shown that seeds persist for a few years or less (Alvarez-Buylla et al. 1990, Dalling et al. 1998a), some pioneers have seed banks that retain high viability for at least three years even when exposed to predators and decomposer organisms (Hopkins and Graham 1987, Dalling et al. 1997, Murray and Garcia 2002).

The adaptive value of seed persistence is to reduce the impact on plant recruitment of temporal variation in the favorability of habitat conditions (Venable and Brown 1988). In environments where recruitment conditions fluctuate strongly over large spatial scales, seed persistence may provide the only solution to increasing population growth rate, and may be critical to maintaining populations of short-lived or semelparous species. In forests however, interannual variation in the recruitment rates of pioneer species may be as strongly affected by the spatial location of recruitment sites as by the frequency of their occurrence (Dalling et al. 1998b). In these environments therefore, dispersal and seed persistence may have equivalent effects on recruitment success. Nonetheless, selection is expected to favor the evolution of enhanced dispersal as

persistence incurs an added fitness cost resulting from delayed reproduction (Venable and Brown 1988). Furthermore, selection on long-term seed persistence should be weak if adult reproductive lifespan exceeds the interval between which favorable recruitment sites become available (Rees 1994, Thompson 2000).

However, selection in favor of dispersal assumes that life-history trait combinations are unconstrained. Increased dispersal ability may carry a cost associated of reduced establishment success, associated with reduced seed mass, or may impose additional constraints on the range of microsites where establishment can successfully occur (Dalling and Hubbell 2002). Here we provide a first step in describing interspecific variation in seed persistence in pioneer trees, using a carbon dating technique to determine the age of viable seeds extracted from natural soil seed banks.

Study Site and Methods

We measured how long seeds of pioneer species remain viable in the soil beneath seasonally moist lowland tropical forest on Barro Colorado Island (BCI), Panama (9°10'N, 79°51'W). Rainfall on BCI averages 2600 mm/yr rainfall with a pronounced dry season from January to April (Windsor 1990). Seeds were collected in May 2002 from soil cores taken within old-growth forest in the 50 ha forest dynamics plot in the center of the island (Hubbell and Foster 1983). Only surface soil layers (0-3 cm depth) were sampled to ensure that seeds were collected from burial depths from which emergence can successfully occur (Pearson et al. 2002). Seeds were extracted from the soil by wet

sieving, identified to species, and germinated in sand in individual Petri dishes. To increase the probability of encountering ‘old’ seeds, we used plot data to sample locations where reproductive-sized individuals of focal tree species had occurred over the previous 20 years.

We targeted three relatively large-seeded pioneer species shown to retain high viability over two years in seed burial experiments (Dalling et al. 1997). *Croton billbergianus* (Euphorbiaceae), air-dry seed mass 24 mg, is a ballistically-dispersed sub-canopy tree with a median dispersal distance of 2.2 m (Dalling et al. 2002). *Croton* is among the commonest pioneers on BCI with 367 reproductive-sized individuals recorded in the 1995 census of the 50 ha plot. *Trema micrantha* (Celtidaceae), seed mass 3.9 mg, is a bird dispersed canopy tree with year-around seed production. *Trema* is rare in old-growth forest on BCI; 11 reproductive sized individuals were recorded in the 1995 plot census. *Zanthoxylum ekmanii* (Rutaceae), seed mass 11 mg, is the most abundant of four congeneric dioecious canopy trees on BCI; 108 reproductive individuals were recorded in the 1995 plot census. Median dispersal of *Zanthoxylum* spp was estimated at 0.8 m (Dalling et al. 2002). *Zanthoxylum* fruits are semi-dehiscent follicles that lack an apparent dispersal reward for frugivores. Nonetheless, seeds are reportedly dispersed by primates (Hladik and Hladik 1969).

Samples of seed coat material from 32 seeds that germinated (supplementary table) were cleaned and dated using accelerator mass spectrometry (AMS) at the Center for Accelerator Mass Spectrometry, Lawrence Livermore National Laboratory. AMS

yields high precision measurements of $^{14}\text{C}/^{13}\text{C}$ based on very small amounts of carbon (<100 μg) by providing counts of individual isotope atoms (Donahue et al. 1990, Donahue 1995, Moriuchi et al. 2000). Dates of carbon fixation for the samples were determined by regressing the sample F^{14}C value (Reimer et al. 2005) against a long-term atmospheric record for Northern Hemisphere zone 2 (Hua and Barbetti 2005). Seed dates obtained with this technique are conservative because we assume that carbon fixation occurred during the period of declining $^{14}\text{C}/^{13}\text{C}$ (post 1963), rather than ascending $^{14}\text{C}/^{13}\text{C}$ (1953-1963). As an independent check of our ability to accurately date seeds we analyzed fruit and seed wall material from dated herbarium specimens collected and stored on BCI. Specimens used were *Ficus insipida*, *Hamelia axillaris*, *Miconia lonchophylla*, *Ochroma pyramidale*, *Simarouba amara*, *Stemmadenia grandiflora*, *Stenospermaton angustifolium* and *Tabebuia rosea* collected between August 1967 and August 1996.

Results and Discussion

AMS carbon dating of herbarium seed material of known age (Fig. 1) shows that predicted ages for field-collected seeds are likely to be accurate within <2 years. With one exception, predicted seed ages and seed collection dates differed by <18 months. Seeds of *Stenospermaton angustifolium*, an epiphytic aroid collected in 1997, however, were predicted to be >4 years older than observed. The age discrepancy for this species probably indicates long-term storage of fixed carbon in rhizomes used for infrequent reproduction.

Application of AMS dating to field-collected seeds indicates that seeds of the three pioneer species tested are capable of persisting in the soil for decades (Fig. 2). Germinable seeds were recovered from surface soil layers after up to 18 yr (*Zanthoxylum*), 31 yr (*Trema*) and 38 yr (*Croton*) in the soil seed bank. These persistence times are remarkable given the rapid decomposition rates of woody material on the soil surface in tropical forests (Chambers et al. 2000).

The costs of AMS prevents dating of sufficient samples to construct survivorship curves for seeds in the seed bank. Nonetheless, the relatively high viability of seeds isolated from the same soil cores from which seeds were dated suggests that long-term persistence is likely to be common for these species (supplementary table). For *Zanthoxylum*, three seeds dated from one soil core gave consistent ages of 15-16 years; 11 % of the 535 seeds recovered from the core were germinable. For *Trema*, five seeds dated from one core gave ages ranging from 20-31 years; 12 % of 100 seeds recovered from the core were germinable. For *Croton*, a sub-canopy tree with very low estimated fecundity (2.6 seeds/cm² basal area; Dalling et al. 2002), only ten seeds were recovered from the soil, seven of which were germinable.

The seed ages reported here greatly extend the measured duration of seed persistence times for tropical pioneers. Seed burial experiments have shown that seeds of the majority of pioneer species retain high viability for at least two years (Perez-Nasser and Vázquez-Yanes 1986, Hopkins and Graham 1987, Dalling et al. 1997), but longer-term burial experiments have not been reported. Decadal seed persistence however is

consistent with inferred seed persistence times calculated as the ratio of seed bank density to average annual seed rain. Repeated measurements of seed rain and soil seed densities made in a montane forest in Costa Rica yielded a median seed bank to seed rain ratio 8.4 for 23 species (Murray and Garcia 2002). Several species common in the seed bank, including *Trema micrantha*, were not observed in seed rain over three years, while a pioneer shrub *Bocconia frutescens* accumulated a seed bank equivalent to 85 year of seed rain. Estimates of seed persistence based on seed bank to seed rain ratios are sensitive to sampling methods and are unlikely to meet assumptions of spatial homogeneity and temporal equilibrium of seed rain and seed banks (Garwood 1989, Murray and Garcia 2002). Nonetheless, the combination of direct seed dating and inferred seed residence times in the soil strongly suggest that long-term seed persistence is common for tropical pioneer trees and shrubs.

In contrast to field data, population growth models predict that selection for long-term seed persistence should be weak (Murray 1988). This is because seeds that germinate after a long period in the soil contribute less to population growth than those that arrive directly in gaps or germinate soon after dispersal. Long-term seed persistence, however, may arise even if it is not under direct selection. Risk of mortality for seeds is highest shortly after dispersal, when seeds are exposed to predators and pathogens on the soil surface (Estrada and Coates-Estrada 1991, Dalling et al. 1997). Physical and chemical traits that provide sufficient protection for seeds to become incorporated into the soil may incidentally also confer long-term persistence in the soil.

Alternatively, long-term seed persistence may indicate that reproductive models for pioneers require revision. Predictions that long-persistent seeds contribute little to population growth based on population projection models are strongly influenced by parameter estimates for the proportion of seeds that germinate each year (Murray 1985, Murray 1988). Estimates of germination success based on seed sowing experiments in gaps (Murray 1988) are several orders of magnitude greater than observed probabilities of seedling recruitment based on comparisons of seed bank and seedling recruit densities in artificially created gaps (Dalling and Hubbell 2002). Finally, projection models have also assumed that recruitment rates from seed banks are constant over time. Selection for long-term seed persistence may reflect strong inter-annual variation in the timing and frequency of gap formation.

Although species included in this study have similar seed mass they differ in other important respects. *Croton* is slow growing, has short-distance dispersal, and is capable of regenerating in a wide range of gap sizes (Pearson et al. 2003). *Zanthoxylum* appears to have similarly short-distance dispersal, but is among the fastest growing species in the forest (Condit et al. 1993). *Trema* is rare on BCI and is restricted to very large treefall gaps that occur infrequently in old growth forest (Brokaw 1987, Pearson et al. 2003). Thus long-term seed persistence does not appear to be restricted to pioneers that share a single combination of life-history traits but instead is likely to be widespread among species in this functional group.

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229 **Literature cited**

230 Alvarez-Buylla, E.R., and M. Martínez-Ramos. 1990. Seed bank versus seed rain in the
231 regeneration of a tropical pioneer tree. *Oecologia* 84:314-325.

232 Brokaw, N.V.L. 1987. Gap-phase regeneration of three pioneer tree species in a tropical
233 forest. *Journal of Ecology* 75:9-19.

234 Chambers, J.Q., N. Higuchi, J.P. Schimel, L.V. Ferreira, and J.M. Melack. 2000.
235 Decomposition and carbon cycling of dead trees in tropical forests of the central
236 Amazon. *Oecologia* 122:380-388.

237 Condit R., S.P. Hubbell, and R.B. Foster. 1993. Identifying fast-growing native trees
238 from the neotropics using data from a large, permanent census plot. *Forest*
239 *Ecology and Management* 62:123-143.

240 Coomes D.A., and P.J. Grubb. 2003. Colonization, tolerance, competition and seed-size
241 variation within functional groups. *Trends in Ecology and Evolution* 18:283-291.

242 Dalling, J.W., M.D. Swaine, and N.C. Garwood. 1997. Soil seed bank community
243 dynamics in seasonally moist lowland forest, Panama. *Journal of Tropical Ecology*
244 13:659-680.

245 Dalling, J.W., M.D. Swaine, and N.C. Garwood. 1998a. Dispersal patterns and seed bank
246 dynamics of pioneer tree species in moist tropical forest, Panama. *Ecology*
247 79:564-578.

- 248 Dalling J.W., S.P. Hubbell, and K. Silvera. 1998b. Seed dispersal, seedling emergence
 249 and gap partitioning in gap-dependent tropical tree species. *Journal of Ecology*
 250 86:674-689.
- 251 Dalling, J.W., and S.P. Hubbell. 2002. Seed size, growth rate and gap microsite
 252 conditions as determinants of recruitment success for pioneer species. *Journal of*
 253 *Ecology* 90:557-568.
- 254 Dalling J.W., H.C. Muller-Landau, S.J. Wright, and S.P. Hubbell. 2002. Role of dispersal
 255 in the recruitment limitation of neotropical pioneer species. *Journal of Ecology*
 256 90:714-727.
- 257 Dalling, J.W., K. Winter and S.P. Hubbell. 2004. Variation in growth responses of
 258 neotropical pioneer species to simulated gaps. *Functional Ecology* 18:725-736.
- 259 Dalling, J.W., and D.F.R.P. Burslem. 2005. Role of life-history and performance trade-
 260 offs in the equalization and differentiation of tropical tree species. Pages 65-88 *in*
 261 D.F.R.P. Burslem, M.A. Pinard, and S.E. Hartley, eds. *Biotic Interactions in the*
 262 *Tropics*. Cambridge University Press, Cambridge.
- 263 Donahue, D. J., T.W. Linick, and A.J.T. Jull. 1990. Isotope-ratio and background
 264 corrections for accelerator mass spectrometry radiocarbon measurements.
 265 *Radiocarbon* 32:135-142.
- 266 Donahue, D. J. 1995. Radiocarbon analysis by accelerator mass spectrometry.
 267 *International Journal of Mass Spectrometry and Ion Processes* 143:235-245.
- 268 Estrada, A., and R. Coates-Estrada. 1991. Howler monkeys (*Alouatta palliata*), dung
 269 beetles (Scarabaeidae) and seed dispersal: ecological interactions in the tropical
 270 rain forest of Los Tuxtlas, Mexico. *Journal of Tropical Ecology* 7:459-474.

- 271 Garwood, N.C. 1989. Tropical soil seed banks. Pages 149-209 *in* M. Leck, V. Parker, and
 272 R. Simpson, eds. Ecology of soil seed banks. Academic Press. San Diego.
- 273 Hartshorn, G. 1990. An overview of neotropical forest dynamics. Pages 585-599 *in* A. H.
 274 Gentry, ed. Four neotropical rainforests. Yale University Press, New Haven,
 275 Connecticut, USA.
- 276 Hladik, A., and C. M. Hladik. 1969. Rapports trophiques entre vegetation et primates
 277 dans la floret de Barro Colorado (Panama). *La Terre et al Vie* 1:25-117.
- 278 Hopkins, M.S., and A.W. Graham. 1987. The viability of seeds of rainforest species after
 279 experimental soil burials under tropical wet lowland forest in north-eastern
 280 Australia. *Australian Journal of Ecology* 12:97-108.
- 281 Hua, Q., and M. Barbetti. 2004. Compilation of tropospheric bomb ^{14}C data for carbon
 282 cycle modeling and age calibration purposes. *Radiocarbon* 46:1273-1298.
- 283 Hubbell, S.P., and R.B. Foster. 1992. Short-term dynamics of a neotropical forest – why
 284 ecological research matters to tropical conservation and management. *Oikos* 63:
 285 48-61.
- 286 Hubbell, S. P., and R.B. Foster. 1983. Diversity of canopy trees in a neotropical forest
 287 and implications for conservation. Pages 25-41 *in* S. Sutton, T. C. Whitmore and
 288 A. Chadwick, eds. Tropical rain forests: Ecology and management. Blackwell
 289 Scientific Publications, Oxford.
- 290 Hubbell, S.P., R.B. Foster, S.T. O’Brien, K.E. Harms, R. Condit, R. B. Wechsler, S.J.
 291 Wright, and S. Loo de Lao. 1999. Light gap disturbances, recruitment limitation,
 292 and tree diversity in a neotropical forest. *Science* 283:554-557.

- 293 Kitajima, K. 1994. Relative importance of photosynthetic traits and allocation patterns as
 294 correlates of seedling shade tolerance of 13 tropical trees. *Oecologia* 98:419-428.
- 295 Molino, J-F., and D. Sabatier. 2001. Tree diversity in a tropical rain forests: A validation
 296 of the intermediate disturbance hypothesis. *Science* 294:1702-1704.
- 297 Moriuchi, K.S., D.L. Venable, C.E. Pake, and T. Lange. 2000. Direct measurement of
 298 the seed bank age structure of a Sonoran Desert annual plant. *Ecology* 81:1133-
 299 1138.
- 300 Murray, B.G. 1985. Population growth as a measure of individual fitness. *Oikos* 44:509-
 301 511.
- 302 Murray, K.G. 1988. Avian seed dispersal of three neotropical gap-dependent plants.
 303 *Ecological Monographs* 58:271-298.
- 304 Murray, K.G., and M. Garcia. 2002. Contributions of seed dispersal to recruitment
 305 limitation in a Costa Rican cloud forest. Pages 323-338 *In* D.J. Levey, W.R. Silva,
 306 and M. Galetti, eds. *Seed dispersal and Frugivory: Ecology, Evolution and*
 307 *Conservation*. CABI, Wallingford
- 308 Pearson, T.R.H., D.F.R.P. Burslem, C.E. Mullins, and J.W. Dalling. 2002. Germination
 309 ecology of neotropical pioneers: Interacting effects of environmental conditions
 310 and seed size. *Ecology* 83:2798-2807.
- 311 Pearson, T.R.H., D.F.R.P. Burslem, R.E. Goeriz, and J.W. Dalling. (2003). Regeneration
 312 niche partitioning in neotropical pioneers: effects of gap size, seasonal drought
 313 and herbivory on growth and survival. *Oecologia* 137:456-465.

- 314 Perez-Nasser, N., and C. Vasquez-Yanes. 1986. Longevity of buried seeds from some
315 tropical rain forest trees and shrubs of Veracruz, Mexico. *Malayan Forester*
316 49:352-356.
- 317 Rees, M. 1994. Delayed germination of seeds: A look at the effects of adult longevity, the
318 timing of reproduction and population age/stage structure. *American Naturalist*
319 144:43-64.
- 320 Reimer, P.J., T.A. Brown, and R.W. Reimer. 2004. Discussion: Reporting and calibration
321 of post-bomb ¹⁴C data. *Radiocarbon* 46:1299-1304.
- 322 Schnitzer, S.A., J.W. Dalling, and W.P. Carson. 2000. The impact of lianas on tree
323 regeneration in tropical forest canopy gaps: evidence for an alternative pathway of
324 gap-phase regeneration. *Journal of Ecology* 88:655-666.
- 325 Swaine, M.D., and T.C. Whitmore. 1988. On the definition of ecological species groups
326 in tropical rain forests. *Plant Ecology* 75:81-86.
- 327 Thompson, K. 2000. The functional ecology of soil seed banks. Pages 215-235 *in* M.
328 Fenner, ed. *Seeds: The ecology of regeneration in plant communities* (2nd edition)
329 CABI, Wallingford.
- 330 Venable, D.S., and J.S. Brown. 1988. The selective interactions of dispersal, dormancy
331 and seed size as adaptations for reducing risk in variable environments. *American*
332 *Naturalist* 131:360-384.
- 333 Whitmore, T.C. 1983. Secondary succession from seed in tropical rain forests. *Forestry*
334 *Abstracts* 44:767-779.

- 335 Windsor, D.M. 1990. Climate and moisture variability and tropical forest: long-term
336 records from Barro Colorado Island, Panama. Smithsonian Institution,
337 Washington D.C.
- 338 Young, T.P., and S.P. Hubbell. 1991. Crown asymmetry, treefalls, and repeat disturbance
339 of broadleaf forest gaps. *Ecology* 72:1464-1471.
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342 Supplementary Table: Seeds of the three focal species *Croton bilbergianus*, *Zanthoxylum*
 343 *ekmannii*, and *Trema micrantha*, were collected from sites beneath living or recently dead
 344 conspecific adults in the BCI 50 ha forest dynamics plot. Tree tag identifies the
 345 individual tree below which trees were collected. Locations of these trees are available
 346 from on-line plot census data (www.ctfs.si.edu/datasets/bci). Dates of tree death are
 347 known to the five year intervals between censuses of the 50-ha plot (Hubbell and Foster
 348 1983). Seed date gives the predicted date of carbon fixation for each seed. The percent of
 349 seeds viable and the total number of seeds recovered is given for each tree location.
 350

| Species | Tree Tag | Date of Tree Death | Seed Date | % Seeds viable (n) |
|--------------------|----------|--------------------|-----------|--------------------|
| <i>Croton</i> | 241614 | 1995-2000 | 1963.9 | 100 (3) |
| <i>Croton</i> | 241614 | 1995-2000 | 1993.4 | - |
| <i>Croton</i> | 241614 | 1995-2000 | 1990.4 | - |
| <i>Croton</i> | 417588 | Alive 2002 | 2000.1 | 33 (3) |
| <i>Croton</i> | 135254 | 1995-2000 | 1992.7 | 100 (2) |
| <i>Croton</i> | 135254 | 1995-2000 | 1984.2 | - |
| <i>Croton</i> | 163343 | 1995-2000 | 1999.1 | 50 (2) |
| <i>Zanthoxylum</i> | 7318 | Alive 2002 | 1999.6 | 4 (472) |
| <i>Zanthoxylum</i> | 7318 | Alive 2002 | 2000.9 | - |
| <i>Zanthoxylum</i> | 7318 | Alive 2002 | 2000.8 | - |
| <i>Zanthoxylum</i> | 4240 | Alive 2002 | 1999.6 | 3 (402) |
| <i>Zanthoxylum</i> | 4240 | Alive 2002 | 2000.1 | - |
| <i>Zanthoxylum</i> | 4240 | Alive 2002 | 1995.4 | - |

| | | | | |
|--------------------|------|------------|--------|----------|
| <i>Zanthoxylum</i> | 4240 | Alive 2002 | 1999.1 | - |
| <i>Zanthoxylum</i> | 3846 | 1995-2000 | 1987.7 | 11 (535) |
| <i>Zanthoxylum</i> | 3846 | 1995-2000 | 1986.5 | - |
| <i>Zanthoxylum</i> | 3846 | 1995-2000 | 1987.6 | - |
| <i>Zanthoxylum</i> | 3846 | 1995-2000 | 1986.1 | - |
| <i>Zanthoxylum</i> | 6263 | 1982-1985 | 2003.5 | 22 (54) |
| <i>Zanthoxylum</i> | 6263 | 1982-1985 | 2001.6 | - |
| <i>Zanthoxylum</i> | 6263 | 1982-1985 | 2002.8 | - |
| <i>Zanthoxylum</i> | 6263 | 1982-1985 | 2001.4 | - |
| <i>Trema</i> | 3085 | 1985-1990 | 1981.4 | 12 (100) |
| <i>Trema</i> | 3085 | 1985-1990 | 1971.6 | - |
| <i>Trema</i> | 3085 | 1985-1990 | 1977.4 | - |
| <i>Trema</i> | 3085 | 1985-1990 | 1980.8 | - |
| <i>Trema</i> | 3085 | 1985-1990 | 1982.3 | - |
| <i>Trema</i> | 4629 | 1985-1990 | 1985.3 | 8 (51) |
| <i>Trema</i> | 4629 | 1985-1990 | 1982.1 | - |
| <i>Trema</i> | 4629 | 1985-1990 | 1976.9 | - |
| <i>Trema</i> | 6102 | 1982-1985 | 1973.9 | 66 (9) |
| <i>Trema</i> | 6102 | 1982-1985 | 1997.4 | - |

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Figure Legends

Figure 1. Correlation of date of seed production derived from ^{14}C dating with date of seed collection for eight liana and tree species sampled from herbarium sheets. Correlation $r=0.99$. Error bars are 95 % confidence intervals for the regression of the isotopic ^{14}C ratio determined by mass spectrometry and the atmospheric F^{14}C record for Northern Hemisphere Zone 2 (Hua and Barbetti 2004, Reimer et al. 2004).

Figure 2. Atmospheric record of F^{14}C for Northern Hemisphere Zone 2 (grey points), and predicted dates and their 95 % confidence intervals for the production of individual seeds of a) *Croton bilbergianus*, b) *Trema micrantha* and c) *Zanthoxylum ekmanii* recovered from surface soils in lowland wet tropical forest. Dates for seeds were determined from regressions of seed F^{14}C s with the atmospheric record.

Fig. 1

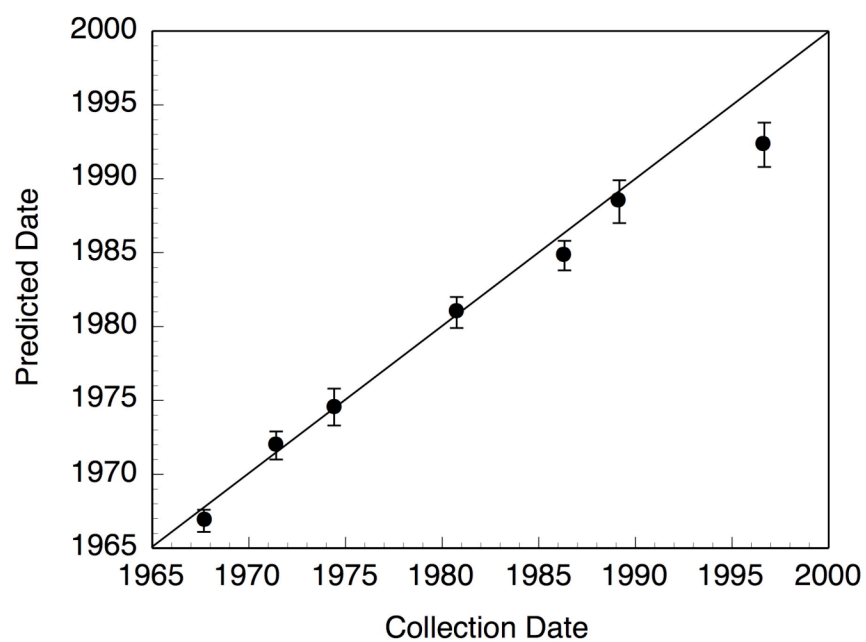
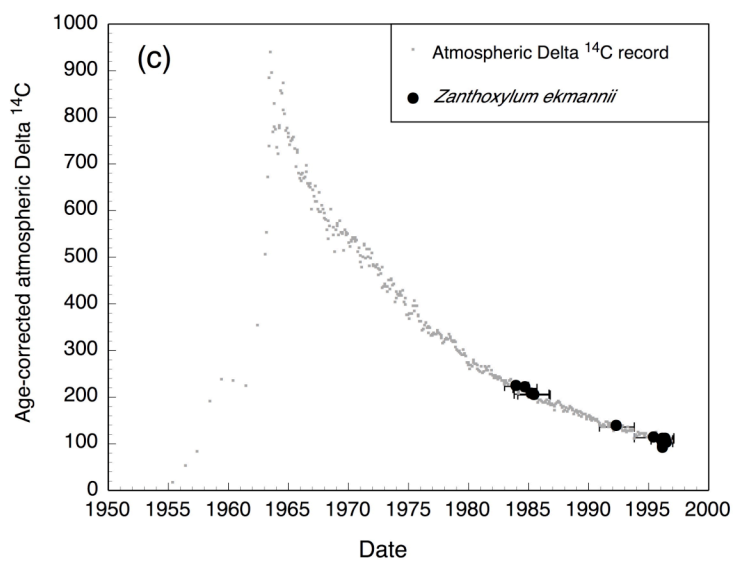
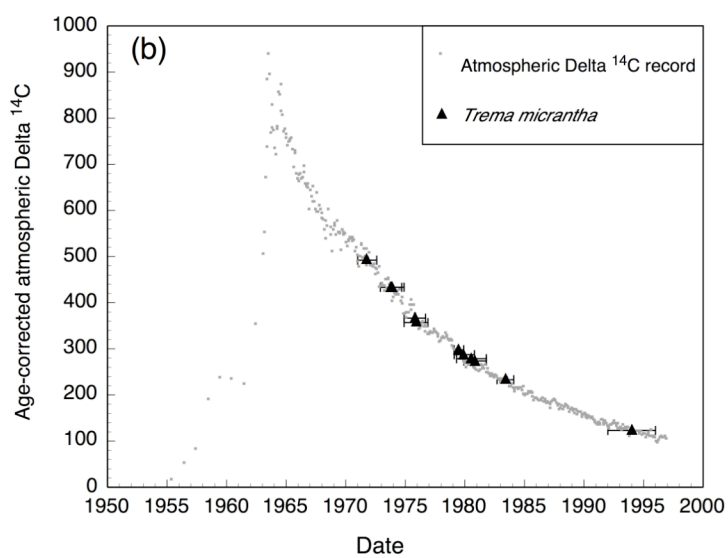
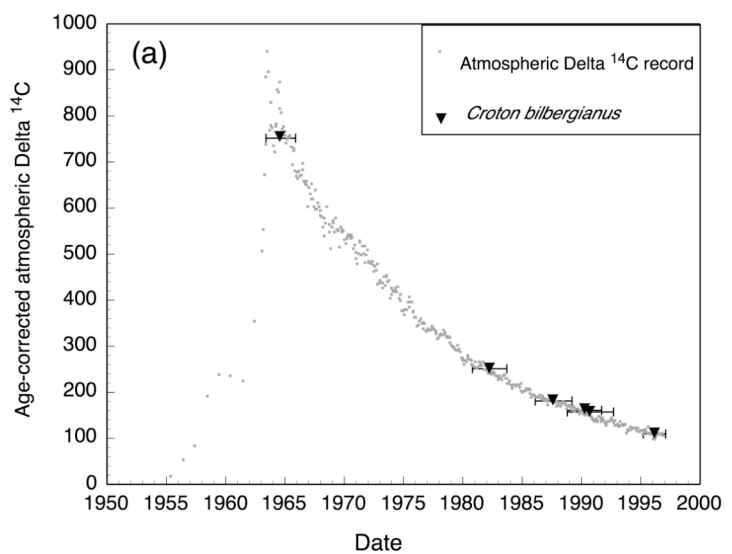


Figure 2



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